# Rheological and mechanical properties of recycled polyethylene films contaminated by biopolymer Gere D., Czigány T.

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1	Rheological and mechanical properties of recycled polyethylene films contaminated by
2	biopolymer
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11	
12	Abstract. Nowadays, with the increasing amount of biopolymers used, it can be expected that
13	biodegradable polymers (e. g. PLA, PBAT) may appear in the petrol-based polymer waste
14	stream. However, their impact on the recycling processes is not known yet; moreover, the
15	properties of the products made from contaminated polymer blends are not easily predictable.
16	Therefore, our goal was to investigate the rheological and mechanical properties of synthetic
17	and biopolymer compounds. We made different compounds from regranulates of mixed
18	polyethylene film waste and original polylactic acid (PLA) by extruison, and injection molded
19	specimens from the compounds. We investigated the rheological properties of the regranulates,
20	and the mechanical properties of the samples. When PLA was added, the viscosity and specific
21	volume of all the blends decreased, and mechanical properties (tensile strength, modulus, and
22	impact strength) changed significantly. Young's modulus increased, while elongation at break
23	and impact strength decreased with the increase of the weight fraction of PLA.
24	
25	Keywords: recycled polyethylene; polylactic acid; viscosity; pvT diagrams; mechanical
26	properties
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29	1. Introduction
30	Nowadays most plastic (about 40% of all plastic) is processed by the packaging sector (Gilbert,
31	2017). The polymers most commonly used for packaging are low density polyethylene (LDPE),
32	linear low density polyethylene (LLDPE), high density polyethylene (HDPE), polypropylene
33	(PP), poly(ethylene terephthalate) (PET), polystyrene (PS) and polyvinyl chloride (PVC)
34	(Gilbert, 2017). The main fields of application of LDPE and LLDPE are reusable bags, trays

and containers, agricultural films and food packaging films (Andreoni et al., 2015). According
to Plastics News, PET (60%) and PE (35%) accounted for most recycled plastic, due to the
short lifetime of packaging materials (Biron, 2017). A wide variety of new products can be
produced from plastic waste, therefore it has to be collected and recycled (Andreoni et al., 2015,
Sharma and Bansal, 2016).

6 However, lately there has been great demand for packaging made from bioplastics. In 2015, the 7 global production of petroleum-based plastics exceeded 300 million tonnes (Emadian et al., 8 2017). Nevertheless, in 2014, 1.7 million tonnes of bioplastics were manufactured in the world 9 with almost 75% of the volume destined for the packaging market (Biron, 2017; Emadian et 10 al., 2017). According to European Bioplastics, bioplastics have three main groups: biobased, 11 non-biodegradable materials (bio-PE, bio-PET); biobased and biodegradable materials (PLA, 12 starch blends); and fossil-based and biodegradable materials (poly(buthylene-adipate-co-13 terephthalate) (PBAT)). Shopping bags, mulch films, and yoghurt cans can be made from PLA 14 (Tábi et al., 2010). However, PLA (4.75 \$/kg) is more expensive than PE (1.9 \$/kg) or PET (1.6 15 \$/kg) (Biron, 2017; Emadian et al., 2017).

16 The potential risk of contamination of the collected conventional plastics and the cost of 17 separation have an impact on the quality and processing of recycled materials. For example, 18 PLA and PET cannot be easily or cheaply sorted by sight or by separation methods based on 19 density because their densities are similar (Gent et al., 2009). Therefore, hybrid bioplastics 20 (blends of bio-based and petroleum-based polymers) appeared. While some biopolymers may 21 have little effect on recycled HDPE, the presence of biopolymers can significantly modify the 22 properties of PET (Cornell, 2007; La Mantia et al., 2012). The presence of biopolymers in post-23 consumer plastic waste makes the recycling of bioplastic blends complicated (Soroudi and 24 Jakubowicz, 2013).

25 Yarahmadi et al., 2016 investigated the effects of reprocessing and recycling on PLA and a 26 PLA/HDPE blend. They used a commercially available (RTP Company) PLA/HDPE blend 27 which contained 39% PLA. To reduce compatibility problems, they used glycidyl methacrylate-28 grafted PE as compatibilizer. The materials were evaluated after a varying number of processing 29 steps (up to 6 consecutive extrusions). Multiple processing of the blend did not significantly 30 change the elastic modulus, but elongation at break was significantly affected by the number of 31 extrusions. Two basic mechanisms can affect elongation at break. One is the degradation of one 32 or both polymers, which decreases elongation at break (La Mantia, 1996). The second is the 33 increased dispersion of the dispersed phase, which means smaller domain size and lower 34 interfacial tension, which causes an increase in elongation at break (Fortelný et al., 2011). The

1 results of Yarahmadi et al., 2016 indicated that multiple processing of a PLA/HDPE blend 2 caused increased dispersion of the PLA in the HDPE, and thus increased elongation at break. 3 Using SEM, they observed the cryofractured surfaces of the samples and found that the size of 4 the particles and holes in the blend extruded once were in the range of  $1.0-1.4 \,\mu\text{m}$ , whereas in 5 the blend extruded six times, the range was 0.5-0.9 µm. The DSC thermograms of the blends 6 showed that the components of the blend were not miscible. The blending of PLA with HDPE 7 not only caused the crystallization of the PLA, but also led to a significant decrease in its cold 8 crystallization temperature and a significant increase in its melting peak temperature. HDPE 9 acted as a nucleating agent and lowered the surface free energy barrier toward nucleation and 10 thus initiated crystallization.

PLA has several advantages: it is recyclable, compostable and can be produced from renewable resources. In addition, it has good stiffness and strength, but it has some disadvantages too, such as moisture sensitivity, poor impact resistance and low heat deflection temperature. There are numerous methods to modify its properties, such as plasticization, copolymerization, blending, or producing composites (As'habi et al., 2013; Chow et al., 2016; Graupner et al., 2016; Imre et al., 2014; Tábi et al., 2016).

17 Abdolrasouli et al., 2015 studied virgin PLA and PLA/PE blends which contained 10%, 20% 18 and 30% PE. They investigated the effects of PE content on morphology and the linear 19 viscoelastic properties of the melt. The minor phase was a linear low density polyethylene co-20 polymer (PE) with butane as co-monomer. The samples were prepared with an internal mixer 21 (Brabender Plasticorder W50) at a temperature of 180°C. They studied the linear viscoelastic 22 behavior of the samples in the melt with a rheometric mechanical spectrometer (Paar Physica 23 UDS 200). They performed all the measurements at 180°C in a parallel plate fixture. When the 24 concentration of PE increased from 10% to 20%, the diameter of the dispersed phase increased 25 and the relaxation process became longer, leading to an increase in the storage modulus. The 26 blend containing 30% of PE showed higher elasticity than both PLA and the PE homo-polymer 27 in the low frequencies. Complex viscosity increased with the increase of the weight fraction of 28 PE, even if this effect was not significant in the higher (100-1000 1/s) shear rate range.

Only a few publications examine how the presence of biopolymers affects the properties of conventional petrol-based polymers in their blends. However, according to a recent study of Plastics Recyclers Europe, 2017 it is a considerable problem in Southern Europe because there, in mixed film waste, the share of biodegradable plastics (e.g. starch, PLA, polybutylene adipate terephthalate (PBAT)) is higher than in Northern Europe. Film production with recycled plastics from Northern Europe worked normally, but holes and peaks occurred regularly when 1 recycled plastics from Southern Europe were used. The laboratory tests showed that most of

- the degradation is caused by substances which are used for the production of biodegradablepolymers.
- 4 Nowadays, when mixed polyethylene film waste is recycled, it is a great problem that small 5 amounts (0.5-1.5%) of bioplastics are often mixed into synthetic polymers. The changes in the 6 properties of a product are not easily predictable, but they are very important to know for the 7 manufacturing of products. The novelty of this manuscript, compared to other publications, is 8 that our goal is not to develop a new blend to improve the properties of PLA or PE, but to 9 analyze the influence of bioplastic contamination in the petrol-based polymer waste stream on the recycling process. Therefore, our goal was to investigate the rheological and mechanical 10 11 properties of synthetic and biopolymer compounds.
- 12

#### 13 2. Experimental

# 14 **2.1. Materials**

- LDPE, PP, HDPE and PET are the most common types of plastic processed by the packaging industry and the composition of the plastic fraction of municipal solid waste by polymer type was LDPE (27.9%), HDPE (14%), PP (17.5%) and PET (11.9%) in Europe (Dahlbo et al., 2018;
- 18 PlasticsEurope, 2016). According to Biron, 2017, PET (60%) and PE (35%) account for most
- 19 recycled plastic, PP accounting for only 2%. In 2014, PLA was the most used biodegradable
- 20 bioplastic (Emadian et al., 2017), therefore there is a chance that PE and PLA are mixed in the
- 21 polymer waste stream.
- 22 We used regranulates of mixed polyethylene film waste (MPE) (MFR (190°C, 2.16 kg): 0.56
- 23 g/10 min, density: 0.932 g/cm<sup>3</sup>), supplied by Jász-Plasztik Ltd. (Hungary). It is a post-industrial
- 24 mixed film waste, which typically contains 50% HDPE and 50% LDPE. We also used virgin
- 25 PLA type Ingeo 3100HP (MFR (210°C, 2.16 kg): 24 g/10 min, density: 1.24 g/cm<sup>3</sup>), supplied
- 26 by NatureWorks LLC. (USA).
- 27

## 28 **2.2. Material preparation and processing**

Six different MPE-PLA compounds (regranulates) (100/0, 98/2, 90/10, 75/25, 50/50, 0/100
wt/wt%) were produced with a Labtech Scientific LTE 26-48 co-rotating twin-screw extruder.
Before extrusion, PLA was dried at 100°C in a Memmert UF1060 hot air drying oven for 90
minutes. The screw diameter of the extruder was 26 mm and the length/diameter (L/D) ratio
was 48. The temperature profile of the extruder (from intake zone to die) was 190-195-195-

1 200-200-205-210-215-215-220-220-215°C. The melt pressure was 50-55 bar and the

2 throughput was 8 kg/h.

Before injection molding, the compounds were dried at 100°C in a Memmert UF1060 hot air
drying oven for 3 hours. The injection molded samples were made with an Engel ES 200/45
HL-V injection molding machine. Injection rate was 40 cm<sup>3</sup>/s, holding pressure was 800 bar,
holding time was 20 s, residual cooling time was 30 s, and melt and mold temperature were
225°C and 25°C, respectively. The mechanical tests were carried out on ISO standard dumbbell
shaped tensile specimens.

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#### 10 **2.3. Methods**

11 Before the rheological tests, the samples were dried at 100°C in a Memmert UNE 200 hot air 12 drying oven for 3 hours. The shear viscosity of the regranulates was measured with an Instron 13 Ceast SR20 capillary rheometer. The testing temperature in the capillary rheometer was set to 14 190°C. We carried out the measurements in the 100-20000 1/s shear rate range, using four 15 different capillaries:  $L_1 = 5 \text{ mm}$ ,  $D_1 = 1 \text{ mm}$ ;  $L_2 = 10 \text{ mm}$ ,  $D_2 = 1 \text{ mm}$ ;  $L_3 = 20 \text{ mm}$ ,  $D_3 = 1 \text{ mm}$ ; 16  $L_4 = 30$  mm,  $D_4 = 1$  mm. We used Bagley correction to calculate true shear stress and the 17 Rabinowitsch analysis to calculate true shear rate. The specific volume (pvT diagram) of the 18 regranulates was measured with an Instron Ceast SR20 capillary rheometer. We determined the 19 specific volume with the direct method at 6 different pressures (50-1000 bar) and at 22 different 20 temperatures ranging from 230°C to 50°C (Carreau et al., 1997). We tested all the samples at 21 least 8 times; in every sample standard deviation was within 5% (therefore error bars are not 22 visible on the figures).

Tensile tests were carried out on an Instron 3369 universal testing machine. We calculated the tensile modulus between 0.05% and 0.25% strain with a crosshead speed of 1 mm/min, and determined tensile strength and elongation at break with a crosshead speed of 10 mm/min. The measurements were performed on ISO 527-2/1B dumbbell shaped specimens with an overall length of 150 mm and a cross-section of 4 mm  $\times$  10 mm. In every measurement, we tested at least 5 specimens, and calculated the average value and standard deviation.

29 Impact strength was determined with the Izod impact test on a Zwick 5113.100/01 impact tester,

30 with pendulums of 2.75 J and 22 J. The measurements were performed on 2 mm notched ISO

31 180/A dumbbell shaped specimens with a length of 80 mm and a cross-section of 4 mm  $\times$  10

32 mm. The tests were performed at room temperature and at a relative humidity of 50%. In every

- 33 measurement we tested at least 10 specimens, and calculated the average and standard
- 34 deviation.

1 After drying the samples, we measured the residual moisture content with an aboni Hydrotracer 2 FLV water content analyzer. The absolute water content of the material was determined with a 3 chemical method, not based on weight loss during heating. The instrument warmed the PLA 4 granules to 140°C, therefore the water evaporated. The gaseous water reacted with the calcium 5 hydride reagent in a cooled area and hydrogen was formed. A gas sensor detected the hydrogen 6 concentration within the reactor and separate sensors detected the moisture of the ambient air, 7 which came into the reactor at the beginning of the measurement. The software considered these 8 in the calculation of the water content of the PLA granules.

9 The morphology of the blends was studied with a light microscope (Zeiss AXIO Imager M2m)

10 and a scanning electron microscope (Zeiss EVO MA 10). The SEM micrographs were taken

11 from cryo-fractured surfaces.

The core and the skin layer of the injection molded samples were investigated with a Bruker
Tensor 37 Fourier Transform Infrared Spectroscopy (FTIR) equipment (the resolution of the

- 14 instrument is better than  $0.6 \text{ cm}^{-1}$ ).
- 15

## 16 **2.4. Processing experiments**

In the experiments, MPE was mixed with original PLA (instead of PLA waste), because we did
not want to increase the number of influencing factors, we just wanted to discover the effect of
biodegradable polymer "contamination".

20 In the case of PLA, according of the technical data sheet, a moisture content of less than 250 21 ppm (0.025%) is recommended before processing, to prevent hydrolytic degradation. In order 22 to model industrial conditions (PE flakes are usually dried for a short time after the 23 washing/cleaning process), we dried PLA for a shorter time and it had a moisture content of 24 457 ppm before extrusion. Modern film recycling systems (Erema, Starlinger) can process 25 materials (PE, PP) with up to 12% residual moisture. Therefore, drying temperature and/or time 26 can be reduced, and energy consumption, as well as production costs and CO<sub>2</sub> emission can be 27 decreased. However, in the case of hydrophilic biopolymer (PLA) contaminants mixed in 28 synthetic polymers, hydrolytic degradation probably occurs. Melt temperature was 20°C higher 29 than the recommended value for PLA because of the processing melt temperature of MPE, therefore thermal degradation probably occurred. After extrusion the different regranulates 30 31 looked like typical regranulates.

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#### 1 **3. Results and discussion**

## 2 **3.1. Rheological properties of the regranulates**

3 Figure 1 shows the shear viscosity curves of MPE and PLA before and after extrusion at 190°C. 4 Extrusion did not affect the viscosity of MPE, but after extrusion, the viscosity of PLA 5 decreased. In the case of PLA, this decrease was due to degradation reactions such as thermal 6 unzipping degradation, oxidative chain scission, inter- or intramolecular transesterification and 7 hydrolysis in the presence of moisture. Residual moisture after the drying process attacks the ester linkages, causing chain scission and a decrease in molecular weight, and as a result, 8 9 viscosity decreases. Several researchers (Harris and Lee, 2012; Jamshidi et al., 1988; 10 Yarahmadi et al., 2016) came to similar conclusions about the degradation of PLA. Degradation 11 under processing takes place not only in sensitive polymers, such as PET or PLA, but also in 12 polymers that are relatively resistant to degradation, for example PE. In the case of PE, both 13 degradation and cross-linking occur after multiple processing, which can increase melt 14 viscosity. This is also supported by the results of Cornell, 2007.





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**Figure 1.** Viscosity curves of MPE and PLA before and after extrusion in the 100-2000 1/s (a) and the 2000-20000 1/s (b) shear rate range

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show negative deviation in rheological properties (e.g. viscosity, die swell). Therefore, the two
components were probably incompatible in the melt at least in the concentrations investigated
here. Hamad et al., 2012 found similar results for PLA/LDPE blends in the 0.1-200 1/s shear
rate range. They also observed that the viscosity of the blends indicates negative deviation
according to the linear rule of mixtures.





Figure 2. Viscosity curves of the different MPE-PLA regranulates in the 100-2000 1/s (a) and
 the 2000-20000 1/s (b) shear rate range

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Figure 3 shows the pvT curves of the extruded MPE and PLA at 5 different pressures. The specific volume decreased with the increase of pressure. The crystallization of MPE started at ~125°C, whereas the crystallization of PLA started at ~135°C. Shrinkage (the decrease of specific volume) was 17% for MPE and 15% for PLA between 230 °C and 50 °C at 50 bar.



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Figure 3. pvT curves of the extruded MPE (a) and PLA (b) at different pressures

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Figure 4 shows the pvT curves of the different regranulates. The melt pressure for film blowing is about 200-300 bar in industrial extruders, therefore we analyzed the pvT curves at 250 bar. Specific volume decreased with the increase of the weight fraction of PLA but this effect was not significant in the 98/2 blend. This decrease was due to the higher density of PLA. The pvT curve of blends can be estimated based on the rule of mixtures if the pvT curves of the components are known.

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Figure 4. pvT curves of the different MPE-PLA regranulates at 250 bar

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# 11 **3.2.** Mechanical properties of the injection molded samples

Figure 5 shows the tensile stress-strain curves of the six different samples. The curves show that MPE was flexible and PLA was rigid. The stress-strain curves of all the blends showed a decrease in elongation, but an increase in tensile strength when PLA was added, but this effect was not significant in the 98/2 blend.



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Figure 5. Tensile stress-strain curves of MPE-PLA samples with different weight fractions

Figure 6 and Table 1 show the mechanical properties of the injection molded samples. The
100/0 and the 98/2 blend had similar mechanical properties. Tensile strength decreased when

1 10 and 25 wt% PLA was added, but doubled when 50 wt% PLA was added. As the weight 2 fraction of PLA was increased from 0% to 50%, the Young's modulus increased from 232 MPa 3 to 1137 MPa, while elongation at break decreased from 99.53 % to 3.08 %. In addition to the presence of a rigid component (PLA), the degradation of PLA and the increased droplet size 4 5 (see Figure 9e-h) could also lead to a decrease in elongation at break when the weight fraction 6 of PLA was increased. The impact strength of the samples greatly decreased when 25 and 50 7 wt% PLA was added, but the 100/0, 98/2, 90/0 samples were not broken, even when the 22 J 8 pendulum was used. As the weight fraction of PLA was increased, tensile strength and the 9 Young's modulus seemed to change along a parabola, while elongation at break decreased sharply and stayed low, and no tendency could be observed for impact strength. These 10 11 mechanical properties of the blends did not change according to the linear rule of mixtures. 12 Similarly to rheological properties, the polymer blends are typically incompatible if the values 13 of Young's modulus, elongation at break and stress at break are below the line of linear rule of 14 mixtures (negative deviation). These results show that PLA and MPE compose an incompatible 15 blend, which is confirmed by the rheological results. The trend of increasing Young's modulus 16 and decreasing elongation at break well matched the results of Hamad et al., 2012, who studied 17 a PLA/LDPE blend.

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Duonoutry	100%	98% MPE	90% MPE +	75% MPE +	50% MPE +	100% PLA
Property	MPE	+ 2% PLA	10% PLA	25% PLA	50% PLA	
Impact	not broken	not broken	not broken	$6.07 \pm 0.28$	2.56 ± 0.11	
Strength						$3.75 \pm 0.13$
– Izod	not broken					
$[kJ/m^2]$						

**Table 1.** Impact strength of the different injection molded MPE-PLA samples

1 2

#### 3 **3.3. Structure**

4 After the tensile tests, it seemed that there was a core part which was flexible and a skin part 5 which behaved more rigidly. To prove this, we analyzed the cross sections of the tensile bars 6 with a light microscope (Figure 7) and an FTIR spectrometer (Figure 8). In the blends which 7 contained 2 and 10 wt% PLA, the dispersed PLA phases were not visible. However, in the case 8 of the 75/25 blend, the dispersed PLA component was visible in the MPE matrix, furthermore, 9 the structure of the 50/50 blend was close to co-continuous. The micrographs were taken from 10 the edge of the cross section, where both components could be found and the components were 11 not clearly layered. The outer surface of the samples was investigated by FTIR. In this analysis the region between 800 and 1800 cm<sup>-1</sup> is interesting, where the active ester and methyl groups 12 13 of PLA are located. A higher intensity in the -C=O ester carbonyl stretch region (~1720-1775 14 cm<sup>-1</sup>) was observed with the increase of the weight fraction of PLA. The peak maximum was a wave number of 1752 cm<sup>-1</sup>; this peak is related to carboxylic end groups formed during the 15 16 hydrolysis of the ester bonds, which can indicate the degradation of PLA. This is supported by 17 Yarahmadi et al., 2016 and Badia et al., 2012, who observed the influence of multiple 18 processing of neat PLA and PLA blends on degradation. In Figure 8, at the same time, an 19 increase in signal intensity can be observed at the characteristic PLA peak at 1266 cm<sup>-1</sup> (-CO 20 bend) and the peak of 1086 cm<sup>-1</sup> with the increase of the weight fraction of PLA, which was ascribed to symmetric stretching of the -C-O- band of ester groups. Figure 8 shows that the 21 symmetric deformation of the methyl –CH<sub>3</sub> group occurred at 1381 cm<sup>-1</sup> and 1266 cm<sup>-1</sup>, when 22 23 PLA content was increased from 10% to 50%. The changes in these peaks were also observed 24 by Ndazi and Karlsson, 2011, when they investigated the hydrolytic degradation of PLA and by Vasile et al., 2017, when they studied PLA bionanocomposites for food packing. In Figure 25 26 8, the spectra demonstrated the presence of -CH<sub>3</sub>, -CH<sub>2</sub> and C-H groups of highly aliphatic components in the 2800-3000 cm<sup>-1</sup> wavenumber range, due to the MPE component. A decrease 27

- in signal intensity can be observed at the peak of 2848 cm<sup>-1</sup> and the peak of 2916 cm<sup>-1</sup> with the
  increase of the weight fraction of PLA, which was ascribed to the stretching of the –C–H bond
  of the alkanes. This is supported by Das and Tiwari, 2018 and Miandad et al., 2017, who
- 4 observed the pyrolysis of plastic waste.
- 5



- 6 **Figure 7.** Changes in the morphology of MPE/PLA blends (a) 25 wt% PLA, b) 50 wt% PLA)
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Figure 8. FTIR curves of the different MPE-PLA blends in the 500-3000 cm<sup>-1</sup> wavenumber
 range

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We also studied the structure of the blends by SEM (Figure 9). The micrographs were taken from cryo-fractured surfaces. In the case of the 98/2 blend, the dispersed PLA phase formed

1 only a few tenths of a micron droplets (droplet distribution) and then in the 90/10 blend the 2 divided droplets grew a few microns in size. In the blend that contained 25 wt% PLA, the 3 extended dispersed phases were deformed and the spherical droplets were converted to an 4 elongated fibrous shape. Finally, in the 50/50 blend a co-continuous phase formed. These results 5 are consistent with the results of light microscopy. In SEM micrographs, a typical morphology 6 of immiscible polymer blends could be seen if one of the components had a well-defined 7 spherical shape, very broad size distribution and a smooth fracture surface. These observations 8 are also confirmed by rheological and mechanical test results. Kim et al., 2004 had similar 9 results, when they investigated an 80/20 LDPE/PLA blend. Lu et al., 2016 also observed a cocontinuous phase morphology in PLA/HDPE blends (60/40 and 50/50) and a typical island-sea 10 11 type morphology in PLA/HDPE (40/60 and 20/80) blends.







a)





c)



b)





e)





g)



i)







- **Figure 9.** Changes in the morphology of MPE/PLA blends (a-b) 0 wt% PLA, c-d) 2 wt% PLA, e-f) 10 wt% PLA, g-h) 25 wt% PLA, i-j) 50 wt% PLA, k-l) 100 wt% PLA)
- 2 3

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#### 4 **4.** Conclusions

Nowadays, in the recycling of mixed polyethylene film waste, it is a considerable problem that
small amounts (0.5-1.5%) of bioplastic waste can contaminate synthetic polymer waste.
Developing a method of recycling this post-consumer waste of polymer blends is a major
challenge.

9 We investigated the rheological properties of regranulates and mechanical properties of 10 injection molded samples.

11 The viscosity curves of all the blends showed a decrease of viscosity when PLA was added. 12 The melt viscosities of MPE (100/0) and PLA (0/100) were similar, but the viscosities of the 13 blends did not change according to the linear rule of mixtures. Specific volume decreased with 14 the increase of the weight fraction of PLA, which is due to the higher density of PLA. The 15 stress-strain curves showed that MPE was flexible and PLA was rigid. Tensile strength 16 decreased when 10 and 25 wt% PLA was added, but doubled when 50 wt% PLA was added. 17 As the weight fraction of PLA was increased from 0% to 50%, the Young's modulus increased 18 from 232 MPa to 1137 MPa, while elongation at break decreased from 99.53 % to 3.08 %. The 19 impact strength of the samples greatly decreased when 25 and 50 wt% PLA was added, but the 20 100/0, 98/2, 90/0 samples were not broken, even when a 22 J pendulum was used. In SEM 21 micrographs, we observed a co-continuous phase morphology in the 50/50 blend, and a typical 22 island-sea type morphology in the 98/2, 98/10 and 75/25 MPE/PLA blends. Therefore, the two 23 components are incompatible, at least in the concentrations investigated. 24 Based on the results, it seems that the blend containing 2% PLA can still be used for film

1 than 2 % of PLA would probably result in a lot of faulty products, due to the immiscibility of 2 the components. These faulty products cannot be recycled, they just go to a landfill or are used 3 for energy production. The miscibility of the blends can probably be improved by adding 4 compatibilizers, therefore products with a simple geometry and low mechanical loads (e.g. 5 flowerpot, speed bump etc.) can be produced from these contaminated blends. The applicability 6 and cost-effectiveness of compatibilizers needs further research. Our opinion is that 7 biodegradable plastics should be collected into a separate waste stream and then recycled (if 8 possible) or composted.

9

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## 16 **References**

Abdolrasouli, M. H., Nazockdast, H., Sadeghi, G. M. M., Kaschta, J., 2015: Morphology
development, melt linear viscoelastic properties and crystallinity of
polylactide/polyethylene/organoclay blend nanocomposites. J. Appl. Polym. Sci. 132,
41300.

- Andreoni, V., Saveyn, H. G. M., Eder, P., 2015: Polyethylene recycling: Waste policy scenario
   analysis for the EU-27. J. Environ. Manage. 158, 103-110.
- As'habi, L., Jafari, S. H., Khonakdar, H. A., Boldt, R., Wagenknecht, U., Heinrich, G., 2013:
   Tuning the processability, morphology and biodegradability of clay incorporated
   PLA/LLDPE blends via selective localization of nanoclay induced by melt mixing
   sequence. Express Polym. Lett. 7, 21-39.
- Badia, J.D., Strömberg, E., Karlsson, S., Ribes-Greus, A., 2012: Material valorisation of
  amorphous polylactide. Influence of thermo-mechanical degradation on the morphology,
  segmental dynamics, thermal and mechanical performance. Polym. Degrad. Stabil. 97,
  670-678.
- 31 Biron, M., 2017: Industrial Applications of Renewable Plastics. Elsevier Ltd., Oxford.
- Carreau, P. J., De Kee, D. C. R., Chhabra, R. P., 1997: Rheology of polymeric systems
   principles and applications. Hanser Publishers, Munich.

- Chow, W. S., Leu, Y. Y., Ishak, Z. A. M., 2016: Mechanical, thermal and morphological
   properties of injection molded poly(lactic acid)/calcium carbonate nanocomposites.
   Period. Polytech. Mech. Eng. 60, 15-20.
- 4 Cornell, D. D., 2007: Biopolymers in the Existing Postconsumer Plastics recycling Stream. J.
  5 Polym. Environ. 15, 295-299.
- Dahlbo, H., Poliakova, V., Mylläri, V., Sahimaa, O., Anderson, R., 2018: Recycling potential
  of post-consumer plastic packaging waste in Finland. Waste Manage. 71, 52-61.
- 8 Das, P., Tiwari, P., 2018: Valorization of packaging plastic waste by slow pyrolysis. Resour.
  9 Conserv. Recycl. 128, 69-77.
- Emadian, S. M., Onay, T. T., Demirel, B., 2017: Biodegradation of bioplastics in natural
  environments. Waste Manage. 59, 526-536.
- Gent, M. R., Menendez, M., Torano, J., Diego I., 2009: Recycling of plastic waste by density
   separation: prospects for optimization. Waste Manage. Res. 27, 175-187.
- 14 Gilbert, M., 2017: Brydson's Plastics Materials. Elsevier Ltd., Oxford.
- Graupner, N., Albrecht, K., Ziegmann, G., Enzler, H., Müssig, J., 2016: Influence of
  reprocessing on fibre length distribution, tensile strength and impact strength of injection
  moulded cellulose fibre-reinforced polylactide (PLA) composites. Express Polym. Lett.
  10, 647-663.
- Hamad, K., Kaseem, M., Deri, F., 2012: Poly(lactic acid)/low density polyethylene polymer
  blends: preparation and characterization. Asia-Pac. J. Chem. Eng. 7, 310-316.
- Harris, A.M., Lee, E.C., 2012: Durability of polylactide-based polymer blends for injectionmolded applications. J. Appl. Polym. Sci. 128, 2136-2144.
- Imre, B., Renner, K., Pukánszky, B., 2014: Interactions, structure and properties in poly(lactic
   acid)/thermoplastic polymer blends. Express Polym. Lett. 8, 2-14.
- Jamshidi, K., Hyon, S.-H., Ikada Y., 1988: Thermal characterization of polylactides. Polymer.
  29, 2229-2234.
- Kim, Y.F., Choi, C.N., Kim, Y.D., Lee, K.Y., Lee, M.S., 2004: Compatibilization of immiscible
   poly(l-lactide) and low density polyethylene blends. Fibers Polym. 5, 270–274.
- La Mantia, F. P., 1996: Basic concepts on the recycling of homogeneous and heterogeneous
  plastics, in: La Mantia, F. P. (Eds.), Recycling of PVC & mixed plastic waste. ChemTec
  Publisher, Toronto, pp. 63-76.
- La Mantia, F. P., Botta, L., Morreale, M., Scaffaro, R., 2012: Effect of small amounts of
   poly(lactic acid) on the recycling of poly(ethylene terephthalate) bottles. Polym. Degrad.
   Stabil. 97, 21-24.

1	Lu, X., Tang, L., Wang, L., Zhao, J., Li, D., Wu, Z., Xiao, P., 2016: Morphology and properties					
2	of bio-based poly (lactic acid)/high-density polyethylene blends and their glass fiber					
3	reinforced composites. Polym. Test. 54, 90–97.					
4	Miandad, R., Barakat, M.A., Aburiazaiza, A.S., Rehan, M., Ismail, I.M.I., 2017: Effect of					
5	plastic waste types on pyrolysis liquid oil. Int. Biodeterior. Biodegradation 119, 239-252.					
6	Ndazi, B.S., Karlsson, S., 2011: Characterization of hydrolytic degradation of polylactic					
7	acid/rice hulls composites in water at different temperatures. Express Polym. Lett. 5, 119-					
8	131.					
9	PlasticsEurope, 2016: Plastics – the Facts 2016. An analysis of European plastics production,					
10	demand and waste data.					
11	Plastics Recyclers Europe, 2017: Recycled film quality negatively affected by degradable					
12	plastics from Southern Europe, Press release.					
13	<http: news="" recycled-film-quality-negatively-affected-<="" td="" www.plasticsrecyclers.eu=""></http:>					
14	degradable-plastics-southern-europe> (accessed 17.10.23).					
15	Fortelný, I., Hlavatá, D., Horák, Z., Kolařík, J., Sikora, A., 2011: Blending, in: Seidel, A. (Eds.),					
16	Processing and finishing of polymeric materials. John Wiley & Sons, Inc., Hoboken, pp.					
17	86-144.					
18	Sharma, R., Bansal, P. P., 2016: Use of different forms of waste plastic in concrete – a review.					
19	J. Clean. Prod. 112, 473-482.					
20	Soroudi, A., Jakubowicz, I., 2013: Recycling of bioplastics, their blends and biocomposites: A					
21	review. Eur. Polym. J. 49, 2839-2858.					
22	Tábi, T., Hajba, S., Kovács, J. G., 2016: Effect of crystalline forms (α' and α) of poly(laction					
23	acid) on its mechanical, thermo-mechanical, heat deflection temperature and creep					
24	properties. Eur. Polym. J. 82, 232-243.					
25	Tábi, T., Sajó, I. E., Szabó, F., Luyt, A. S., Kovács, J. G., 2010: Crystalline structure of annealed					
26	polylactic acid and its relation to processing. Express Polym. Lett. 4, 659-668.					
27	Vasile, C., Râpă, M., Ștefan, M., Stan, M., Macavei, S., Darie-Nită, R.N., Barbu-Tudoran, L.					
28	Vodnar, D.C., Popa, E.E., Ștefan, R., Borodi, G., Brebu, M., 2017. New PLA/ZnO: Cu/Ag					
29	bionanocomposites for food packaging. Express Polym. Lett. 11, 531-544.					
30	Yarahmadi, N., Jakubowicz, I., Enebro, J., 2016: Polylactic acid and its blends with petroleum					
31	based resins: Effects of reprocessing and recycling on properties. J. Appl. Polym. Sci.					
32	133, 43916.					